1	High Solids Anaerobic Digestion of Municipal Biosolids Pretreated by Thermal
2	Hydrolysis
3	

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6 Abstract

4

5

7 Anaerobic digestion of high solids municipal biosolids is feasible after pre-treatment that 8 reduces the viscosity of biosolids and increases the fraction of soluble organic matter. 9 The process is stable at organic loadings of up to 6 Kg VS/m³ and hydraulic retention 10 times of 9 to 12 days, can achieve more than 60% volatile solids destruction and produce 11 more than 3 $m^{3}CH_{4}/m^{3}$.d, significantly increasing the treatment capacity of existing 12 digesters or, in treatment plants without spare capacity helping postpone, reduce or even 13 avoid costly infrastructure investments. Current pre-treatment technologies need to be 14 optimized for high-solids single-stage mesophilic anaerobic digestion to show stable operation at 6-8d HRT with superior overall VS destruction and gas production, or the 15 16 more complex to operate high-solids temperature-phased anaerobic digestion process 17 must be employed in which further solubilization of organic matter occurs in the 18 thermophilic stage for the mesophilic stage to operate at short hydraulic retention times. 19 20 Keywords: Ammonia toxicity, high-solids anaerobic digestion, mesophilic and 21 temperature-phased anaerobic digestion, thermal hydrolysis pre-treatment

22

23 Introduction

24 Wastewater treatment plants generate biosolids as a by-product. Raw biosolids are a 25 combination of settled sewage solids and excess biomass generated during secondary 26 treatment of wastewater. Prior to their disposal or reuse, regulations require that biosolids 27 be processed to reduce vector attraction and to inactivate pathogens. In the majority of 28 wastewater treatment plants in North America and Europe, fermentation of biosolids in 29 the absence of oxygen or anaerobic digestion (AD) has been the process of choice for this 30 purpose for several decades because it is a relatively simple, stable process that yields 31 methane gas that can be harvested thus recycling a portion of the energy in biosolids as heat and power.

32

33

34 AD is typically carried out at mesophilic temperatures in a single stage with long

35 hydraulic retention time (HRT) of 15 days or more to ensure process stability and

36 consistent performance while reducing cost of implementation and complexity of

- 37 operation. Design HRT is controlled by the particulate nature of biosolids which, to
- 38 become available as substrate to fermentative biomass and be transformed to mainly

39 methane and carbon dioxide, need to undergo a slow, enzyme-mediated hydrolysis step.

40 The particulate nature of biosolids also put a practical limit of around 5% total solids

41 (TS) to the maximum solids concentration that can be processed in AD operations since

42 mixing, heat transfer and pumping all become inefficient and non cost-effective due to

43 high viscosity of the raw biosolids slurry at greater total solids contents.

44

45 Pre-treatment of raw biosolids that reduced their viscosity and increased the availability

46 of organic substrate to the fermentative biomass would make possible the realization of 47 high solids AD (Richards, et al., 1991) which could significantly increase the treatment 48 capacity of a given reactor and in treatment plants without spare capacity help postpone, 49 reduce or even avoid costly infrastructure investments. In addition, high solids AD could 50 make the stabilization and disinfection of biosolids more efficient and less costly to build 51 and operate thus perhaps changing the dynamics from centralized biosolids handling facilities which require expensive piping and pumping networks and are inherently 52 53 environmentally discriminatory to a more decentralized model of biosolids management, 54 or create new opportunities to generate revenue streams by importing external biosolids. 55 56 Technologies that can achieve this goal include: 57 58 Sonication: Application of energy in the ultrasound frequency (typical operating • 59 frequency of 20 KHz) results in the formation, growth and eventual violent 60 collapse of cavitation bubbles which generate intense local heating and high pressure on the liquid-gas interface, high shearing and turbulence in the liquid 61 62 itself and form free radicals (Lehne, et al., 2001). 63 • Ozonation: Application of ozone at concentrations up to 0.5 g/g TS generates 64 hydroxyl radicals, which rapidly oxidize dissolved and particulate organic matter 65 in biosolids (Battimelli, et al., 2003). • Thermal hydrolysis: Application of heat to 100-175 °C for 30 minutes followed 66 by release of reactor pressure that generate shear force that can cause cells to 67 68 rupture and, in combination with the initial pressure/heating step, contributes to 69 the almost complete disintegration of sludge flocs (Haug, et al., 1978). 70 High pressure homogenization: Alkaline pretreatment to weaken cell membranes • 71 followed by intense pressure gradient to burst cells (Stephenson, et al., 2004). 72

73 Although a few reports exist that describe pilot- and full-scale AD installations that 74 incorporate some of these pre-treatment technologies (Kepp, et al., 2000; Skiadas, et al., 75 2005), little or no process data is available for high solids applications. Therefore, the 76 operational limits of high solids AD and whether it may be a stable and reliable process is 77 not known, in particular, in novel configurations such as the temperature-phased 78 anaerobic digestion (TPAD) process which has been shown to enhance biosolids 79 stabilization and disinfection when compared to single-stage mesophilic AD (Schafer, et 80 al., 2003).

81

82 **Objectives**

The objectives of the pilot study were to demonstrate the feasibility of high solids anaerobic digestion of municipal sludge after thermal hydrolysis pre-treatment; to compare the potential for biosolids stabilization and gas production of high-solids, singlestage mesophilic and temperature-phase anaerobic digestion; and to analyze process performance during steady state operation.

88

89 Materials and Methods

90 The high solids anaerobic digestion options investigated included:

- 1. THMAD: A TH (170°C) pre-treatment step with 25 minutes contact time
- 92 followed by mesophilic anaerobic digestion (MAD) with 12-15 days HRT.

- THTPAD: A TH (170°C) pre-treatment step with 25 minutes contact time
 followed by temperature-phased anaerobic digestion (TPAD) with 9-15 days
 combined HRT.
- 96
- 97

Temperature of the MAD process was kept at 36 ± 1 °C for all reactors whereas those of

3. Control: MAD and TPAD without TH pre-treatment with 15 days HRT.

the TPAD process were maintained at 55±1 °C in the thermophilic stages and at 35±1 °C
 in the mesophilic stages.

101

A semi-automated pilot plant (Figure 1) was built that included four egg-shaped digesters
with capacities of 113 and 227 liters. One 113-liter and one 227-liter digesters were
plumbed and operated in series in the TPAD and THTPAD modes. The other 113-liter
digester operated the THMAD process whereas the other 227-liter vessel was the control
MAD digester.

107

All of the digester vessels were made of 0.3 cm thick steel sheet, and included stands.
The stands allowed bolting to the floor and supported the reactor's weight when full. The vessel body consisted of a cylinder and two cones, one cone welded on top (30° slope)
and one at the bottom (60° slope) of the cylinder. All ports were 2 cm half nipples

112 (threaded) soldered onto the body. Vessels withstood 0.5 m WC interior pressure. The

- 113 overflow line ran 30 cm below the liquid level. A liquid overflow collection box was 114 installed on each digester for level control.
- 115

116 Mixing and heating was achieved through external pumping. The outlet port was located 117 at the bottom of the lower cone while two inlet ports were located on top, one on each 118 side of the upper cone. Four heat exchangers were used, one for each digester. Each 119 heat-exchange system had a 40 cm heat exchanger coil made of 1.2 cm ID (1.6 cm OD) 120 stainless steel piping. The diameter of the coil was 50 cm, and the length of the coil 121 piping was 3 m. The heat exchangers were submerged in water baths which temperature 122 was controlled automatically. Flow rates of 230-460 l/h were necessary to ensure 123 sufficient heat transfer to maintain proper mesophilic or thermophilic temperatures in the 124 reactors, which resulted in the pilot digesters being turned over 48 times per day.

125

Each digester had a gas outlet located at the top of the digester. Digester gas leaves the vessels through a U-Leg of 0.6 m to build enough pressure (0.4 m WC) to drive the wettip positive displacement gas meters. Each meter tripped at about 130 ml and can handle 700 liters per day of gas production. The tippers have an accuracy of ± 1 ml. The gas meters operated with a cyclic pressure buildup starting at about 5 cm of WC and built up to about 20 cm of WC and then expelled the gas and returned to the start of the cycle again.

133

134 **Thermal Hydrolysis Prototype**. Thermal hydrolysis was achieved by injecting live

steam to thickened combined primary and waste activated sludge (CPAS) in a 2-liter

reactor to maintain a temperature of 170 °C for 25 minutes, after which pressure was

released instantaneously by transferring sludge into an non-pressurized flash tank and

138 holding it until cooling to less than 100 °C. An electric steam boiler was used to provide

instantaneous and reserve steam capacity. The boiler provided steam at 200°C and 10 bar
 to the thermal hydrolysis unit.

141

Pilot Plant Operation. Digester feed pumps were controlled by automatic timers. Each
digester had its own time switch, which was programmed individually to feed the
digesters at a specified rate. Table 1 summarizes the digester feed rates during steady
state operations. Control MAD and TPAD digesters were fed CPAS from the Southeast
Water Pollution Control Plant (SEP) in San Francisco, CA. THMAD and THTPAD

- 147 digesters were fed dewatered and hydrolyzed CPAS (HCPAS).
- 148

149 The TH and control digesters were allowed to reach steady state for at least 45 days or

three hydraulic residence times before process performance data was collected. For all

151 other tested processes, steady state was assumed after 30 days of operation. However,

152 data was collected since start-up to gain information on the transitional performance of

- 153 the different processes. In addition, the THMAD and THTPAD digesters were operated
- 154 for approximately three weeks at a hydraulic residence time of 20 days to allow for
- 155 acclimation of the bacterial seed from the full-scale mesophilic digesters at the SEP to the 156 new feed quality and rate.
- 157

158 Routine parameters monitored in the influent streams included total and volatile solids

and soluble and total COD whereas parameters monitored in the digested streams

160 included total and volatile solids, soluble and total COD, ammonia, fecal coliforms, total

volatile fatty acids, pH and total alkalinity. All analyses were performed according to

162 Standard Methods (APHA, 1998). In addition, daily gas production and CO₂ content of

- 163 the gas by the ferryte method were recorded.
- 164

165 **Results**

166 Thermal hydrolysis of CPAS resulted in a significant increase in sCOD concentrations 167 (Figure 2) from a range of 8-10 g/L for CPAS to 18-30 g/L for HCPAS at comparable

168 TVS concentrations and the sCOD/COD fraction increased from 0.16 to 0.25 for samples

ranging from 3 to 11% TS. However, although there is a strong positive correlation

170 (r=0.9) between COD and TVS of both CPAS and HCPAS samples, sCOD is not a

171 function of TVS for any of the two streams and sCOD concentrations in HCPAS remain 172 must in the superstant for TVS concentrations of (9) or high m

- 172 practically constant for TVS concentrations of 6% or higher.
- 173

174 Digester loadings for the various processes tested are presented in Figure 3. THMAD and

175 control TPAD thermophilic digesters were similarly loaded at between 4 and 6 Kg

176 VS/m³.d due to high feed concentration (THMAD process) and reduced HRT (TPAD

thermophilic stage). Control MAD and TPAD mesophilic digesters loading rates ranged
 from 1.4 to 2.2 Kg VS/m³.d, typical of established US design practices. For THTPAD

- 1/8 from 1.4 to 2.2 Kg VS/m².d, typical of established US design practices. For THTPAD digesters operated at a combined HRT of 15 days, which combined high solids feed and
- reduced HRT, loading rates were significantly higher, with THTPAD thermophilic
- digester loading rates ranging from 9 to 15 Kg VS/m^3 .d and from 4 to 6 Kg VS/m^3 .d for
- 182 the THTP mesophilic digester.
- 183

184 Despite organic loading rates 3 to 4 times higher than typical US practice, all high solids 185 AD processes tested and TPAD were stable and achieved comparable or higher volatile 186 solids destruction than the control MAD (Figure 4). Table 2 summarizes operational data 187 for all tested processes during steady state operation.

188

189 THMAD digester initially showed a high Total Volatile Acids to Alkalinity (TVA/Alk) 190 ratio of 0.4 compared to a ratio of 0.1 or less for the control MAD but, after four weeks of 191 operation, converged with control MAD ratios suggesting an acclimating period of the 192 anaerobic digestion bacterial consortium to the different feed quality and quantity to the 193 THMAD digester. Similarly, an acclimation period was apparent for the control TPAD 194 process, longer for the thermophilic than the mesophilic stage, likely as a result of the 195 higher operating temperature and loading conditions of the former. In the THTPAD 196 process, the mesophilic digester also presented TVA/Alk ratios around 0.1 during steady 197 state operation, but this parameter ranged between 1.2 and 1.6 for the thermophilic stage 198 in response to the very high loading rates, in particular at 3-day HRT when pH 199 measurements consistently dropped below the neutral point. All other digesters had pH 200 values between 7.1 and 7.9

201

202 During steady state operation at 15 days HRT, VS destruction in the control MAD

203 digester ranged 50-55%, typical of full-scale anaerobic digestion processes while the 204 THMAD process achieved comparable results (Figure 4). When HRT for the THMAD 205 process was reduced to 12 days at the beginning of December 2001, however, VS 206 destruction dropped to around 45% and TVA concentrations increased slightly suggesting

- 207 that the digester did not have much reserve capacity. Control TPAD and THTPAD 208 processes, on the other hand, achieved VS destruction higher than 65% during steady
- 209 state operation at 15 days overall HRT (Figure 4). THTAP digesters overall HRT was 210 reduced to 12 days without observable reduction in VS destruction (Figure 4) suggesting 211 that the process may be resilient to shock loadings, although the feed TS concentration
- 212 during this period was somewhat lower. In control TPAD process, VS destruction in the 213 thermophilic stage ranged between 20 and 30% and the mesophilic stage removed the 214 remainder up to the overall observed 60-70% VS destruction. However, in the THTPAD
- 215 process, the thermophilic stage destroyed between 20 and 40% of incoming VS, the
- 216 mesophilic stage VS removal rate was comparable to that of the control TPAD and, 217 consequently, the overall THTPAD VS destruction was slightly higher that the TPAD 218 process. This observation would suggest great flexibility in the THTPAD process with 219 the first stage performance being dependent upon the feed composition and rate and the

220 second stage being able to adjust to fluctuations in transfer sludge quality for an overall 221 stable and steady performance.

222

223 Although the fraction of influent organic matter destroyed by the various AD processes 224 tested differed by relatively small values, the mechanisms involved appear to be quite

225 different (Figure 5). The control MAD and THMAD processes behaved similarly with

226 approximately 25% of the total COD removed corresponding to the sCOD fraction,

227 suggesting that the increased organic loading to the THMAD digester did not

228 significantly impact the bacterial consortium responsible for MAD beyond concentrating

229 it from 2% to 4% TS and thus approximately doubling the amount of organic matter

- digested to methane and carbon dioxide in THMAD. Therefore, the higher fraction and amount of sCOD present in HCPAS helps develop and support more biomass in the
- amount of sCOD present in HCPAS helps develop and support more biomass in the
 digester to allow for the higher rate of organic loading of the THMAD process as
- 232 compared to the control MAD, but does not increase the rate of destruction of organic
- matter which rate limiting step is still the solubilization of particulate organic matter.
- 235

236 In the control TPAD process, the thermophilic stage was a net generator of sCOD, which 237 gets formed at the expense of particulate COD, but also removed a great deal of total 238 COD and functioned mostly as a methanogenic digester (Table 2) despite the reduced 239 HRT and increased organic loading. The mesophilic stage mimics the performance of the 240 control MAD and THMAD albeit at a reduced HRT (10 d compared to 5 d for control 241 MAD and THMAD), likely due to the increased sCOD fraction generated in the 242 thermophilic stage. The THTPAD thermophilic stage converted particulate into soluble 243 COD at approximately half the rate it removed total COD and thus, when compared to 244 control TPAD, hydrolysis of organic matter was much enhanced to the detriment of 245 methane evolution. On the other hand, about 40% of the total COD removed in the 246 mesophilic stage was in the form of sCOD as a result of the combined increases in sCOD 247 and digester biomass (between 3 and 4% TS), which helps explains the higher degree of 248 overall organic removal when compared with the control MAD process but also the 249 remarkable stability of the THTPAD process at reduced overall HRT of 9 to 12 days (6-8 250 days for the mesophilic stage).

251

252 Gas production fluctuated over the course of the study owing to the fact that analyses 253 were carried out three times a week whereas gas evolution was measured continuously. 254 However, specific gas production for control MAD, THMAD, control TPAD and the 255 THTPAD mesophilic stage was mostly stable and ranged between 800 and 1100 liters per 256 kilogram of VS destroyed (600 to 800 L/Kg VS destroyed for the thermophilic stage of THPAD), an indication that uninhibited digestion was taking place during steady state. 257 258 Methane content in the gas ranged between 64 and 72%. Gas production from the 259 thermophilic digester for THTPAD was lower and ranged between 400 and 600 liters per 260 kilogram of VS destroyed, likely due to ammonia inhibition (Liu and Sung, 2001) and 261 low pH. Methane content in the gas exceeded 60%. Owing to these results and the 262 observed VS loading and destruction rates for the various processes tested, gas 263 production per unit volume of digester capacity for all processes is presented in Table 3. 264 Control TPAD and THTPAD processes increased gas production by 30 to 40% over their control MAD and THMAD counterparts whereas gas production of high solids anaerobic 265 266 digesters (i.e., THMAD and THTPAD processes) exceeded that of the controls by more 267 than 100%.

268

269 Table 4 provides a comparison of biosolids treatment and handling performance with

- 270 MAD (data for 2005) and THTPAD (projected from results from this study) for the SEP,
- which pure-oxygen activated sludge system treats an average daily flow of $10,000 \text{ m}^3/\text{h}$.
- For the SEP, implementation of the THTPAD process would result in a 23% increase in
- 273 VS destruction, a 44% reduction in biosolids hauling and disposal costs and a 30%
- 274 increase in energy generation from biogas.
- 275

276 **Discussion**

277 High solids anaerobic digestion is an attractive option for the design of biosolids handling 278 facilities in highly urbanized areas where land is at a premium, the cost of hauling and 279 reuse of the end product is elevated and energy is a highly valuable commodity. It should 280 also be of great interest as a means to defer costly capital improvements for AD facilities 281 with capacity and, therefore, compliance problems. As shown by the results of this 282 research, it is also technically feasible since typical operating parameters for conventional 283 MAD and TPAD also generally apply to the high solids versions of these processes. 284 However, the mineralization of organic carbon in high solids processes is rendered more 285 efficient, and therefore can be accomplished at higher rates, by the increase in biomass in 286 the digesters and the solubilization of particulate matter during the pre-treatment of raw 287 biosolids. Mass loading of anaerobic digesters can thus be at least tripled from 288 conventional organic loading rates for conventional MAD digesters without negative 289 effects on process performance thus generating significant savings in necessary digester 290 capacity and ancillary equipment (e.g., for pumping, mixing, heating and dewatering 291 biosolids) at the cost of installing raw sludge pre-treatment and adding complexity to the 292 overall process. Similarly, methane gas generation also doubles (in THMAD) or triples 293 (in THTPAD) per unit volume of digester capacity installed in high solids AD when 294 compared to conventional MAD, adding to the perceived advantages of the process.

295

296 Although more complex to operate, THTPAD appears to be superior to THMAD in its 297 ability to stabilize biosolids and produce biogas at an HRT of 15 days, paralleling how 298 conventional MAD and TPAD compare. THTPAD also appears to be more stable and 299 should clearly be the option of choice when high solids anaerobic digestion processes are 300 considered. At an overall HRT of 9 days, the THTPAD was still capable of consistently 301 achieve a 55%VS destruction although the mesophilic stage was being operated close to 302 the washout rate for methanogens of 5 days (Li and Noike, 1992). The extremely high 303 organic loading rate to the thermophilic stage of the THTPAD process (9 to 15 304 kgVS/m³d) appears to shift its performance to that of an acid digester with high 305 concentrations of volatile fatty acids and acidic pH, resulting in a net production of sCOD 306 across this stage. Thus, the THTPAD process would benefit from a segregation of 307 functions with the thermophilic stage mostly carrying out hydrolysis and solubilization of 308 particulate COD and the mesophilic stage doing the bulk of the methanogenesis. The 309 literature abounds with examples (Gosh, et al., 1995; Alexiou and Panter, 2004) 310 describing the benefits of two-phase digestion for biosolids stabilization. The net result is 311 increased destruction of organic matter and thus production of biosolids that are going to 312 be more stable upon disposal or recycling and increased evolution of methane gas.

313

314 Pre-treatment of raw biosolids increased the fraction of sCOD and made possible the 315 higher organic loadings of THMAD by developing and sustaining an increased amount of 316 biomass in the digester after a period of acclimation of a MAD biological seed to the new 317 feed quality and rate by operating the digester at a HRT of 20 days for 3 weeks. As 318 acclimation was gradually achieved, TVA to alkalinity ratios for the THMAD process 319 converged with those of the control MAD digester, but TVA, alkalinity and ammonia 320 concentrations were approximately twice as much in the THMAD digester than in the 321 control MAD as were the total solids. Therefore, contrary to information published

- 322 elsewhere (Kepp, et al., 2000; Jolis, et al., 2002), given the results in Figures 4 (similar
- 323 VS destruction for control MAD and THMAD) and 5 (identical ratios of sCOD/tCOD
- removed for control MAD and THMAD), the above information suggests that THMAD
- is a concentrated MAD process subjected to its limitations with respect to the rate
- limiting step (hydrolysis of particulate matter), extent of VS destruction or minimum
- 327 HRT for stable operation.
- 328

329 This is mostly caused by the leveling off of sCOD content in HCPAS for TS 330 concentrations above 6-7% which results in soluble to total COD ratios for HCPAS with 331 TS content of 7-10% and CPAS with TS content of 2.5-4.5% to similarly range between 332 0.15 and 0.3. This outcome may have been affected by the usage of a low efficiency pilot 333 scale thermal hydrolysis unit as compared to full-scale systems and by employing CPAS 334 for the tests as no benefits in sludge degradability should be expected for pre-treatment of 335 primary sludge (Clark and Nujjoo, 2000) while energy is being consumed. However, 336 reported COD removal results (Kepp, et al., 2000) for the HIAS Wastewater Treatment 337 Plant, Hamar, Norway, which incorporates thermal hydrolysis of combined primary and 338 activated sludge compare well at 52-59% at 17d HRT with observed results of 50-56% at 339 15d HRT in this study, and strongly suggest that thermal hydrolysis pre-treatment is not 340 efficient enough at solubilizing particulate COD to promote stable high solids (e.g., 8-341 12% TS feed) MAD operation at short HRT (less than 10 days) and achieve superior VS 342 destruction and methane gas production as demonstrated for thermally pre-treated WAS 343 (Li and Noike, 1992). Existing pre-treatment technologies need to be optimized or new 344 approaches developed before this goal can be achieved, or the more complex and 345 expensive to operate high solids THTPAD process must be employed as the thermophilic 346 stage produces sufficient sCOD for the mesophilic, methanogenic stage to show stable 347 operation at 6-8d HRT with superior overall VS destruction and gas production. On the 348 other hand, pre-treatment of raw sludge remains critical in reducing the viscosity of the 349 feed that increases the efficiency of pumping, mixing and heat-exchange operations that 350 make the THMAD and THTPAD processes technically feasible.

351

352 THMAD did not show signs of ammonia toxicity although concentrations of total 353 ammonia as nitrogen (TAN) exceeded 2g/L. Thus, methane production was uninhibited 354 and ranged around 1 m³/Kg VS destroyed whereas %VS destruction and VFA/Alk ratio 355 values were comparable to those of control MAD. However, inhibition of gas production 356 may have been observed in the thermophilic stage of the THTPAD process, which may 357 be explained in part by TAN concentrations in excess of 2 g/L. Since the concentrations 358 of free ammonia for the range of observed pH values for the thermophilic stage were 359 below 150 mgNH₃-N/L, an accepted threshold for the onset of toxicity caused by free 360 ammonia (McCarty and McKinney, 1961), the high levels of ammonium ion (Liu and 361 Sung, 2001) most likely contributed to the observed reduced gas production, since the 362 threshold for ammonium ion inhibition to thermophilic cultures has been shown (Lay, et al., 1998) to be lower than that for mesophilic-tolerant bacteria. However, the acidic pH 363 of the thermophilic stage of THTPAD could have also affected gas production, since gas 364 365 production and methane content of gas from acid digesters tend to be lower than that of 366 methane digesters (Gosh, et al., 1995).

368 Conclusions

- 369 After a pre-treatment step that reduces sludge viscosity and increases the fraction of
- 370 soluble COD, high solids anaerobic digestion is stable, produces highly stabilized
- biosolids suitable for reuse and more than doubles biogas generation when compared to
- 372 conventional mesophilic digestion. High solids temperature-phased anaerobic digestion
- 373 was found superior to single-stage mesophilic digestion with respect to process
- 374 flexibility, biosolids stabilization and biogas generation because significant solubilization
- 375 of particulate organic matter occurs in the thermophilic stage that allows the mesophilic
- 376 methanogenic stage to operate efficiently at hydraulic residence times of only 6 days.
- 377 Ammonia inhibition of methanogens was not observed in high solids mesophilic
- anaerobic digestion but some effect was apparent at thermophilic temperatures asmeasured by a decrease in biogas evolution.
- 379 measured by a de 380

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389 References

- 390 Alexiou, I.E. and Panter, K. (2004). A Review of Two-Phase Applications to Define Best
- 391 Practice for the Treatment of Various Waste Streams. Proceedings of *IWA 10th World*
- 392 *Congress on Anaerobic Digestion*, Montreal, Canada, August 29-September 2.
- 393 APHA/AWWA/WEF (1998). Standards Methods for the Evaluation of Water and
- 394 *Wastewater*, 20th Ed. Washington, D.C.
- 395 Battimelli, A., Millet, C., Delgènes, J.P. and Molleta, R. (2003). Anaerobic Digestion of
- Waste Activated Sludge Combined with Ozone Pre-treatment and Recycling. *Wat. Sci. Tech.*, 48(4), 61.
- Clark, P.B. and Nujjoo, I. (2000). Ultrasonic Sludge Pre-treatment for Enhanced Sludge
 Digestion. *Water Environ. Man.*, 14(1), 66.
- 400 Ghosh, S., Buoy, K., Dressel, L., Miller, T., Wilcox, G. and Loos, D. (1995). Pilot- and
- 401 Full-Scale Two-Phase Anaerobic Digestion of Municipal Sludge. *WER*, **67**(2), 206.
- 402 Haug, R.T., Stuckey, D.C., Gosset, I.M. and McCarty, P.L. (1978). Effect of Thermal
- 403 Pretreatment on Digestibility and Dewaterability of Organic Sludges. J.W.P.C.F, 55(1),
 404 73.
- 405 Jolis, D., Jones, B., Zhou, G., Jones, S., Isleta, C. and Solheim, O.E. (2002). Pilot Study
- 406 of Pre-treatment to Mesophilic Anaerobic Digestion for Improved Biosolids Stabilization.
 407 Proceedings of *WEFTEC 02*, Chicago, IL, September 29-October 2.
- 408 Lay, J.J., Li, Y.Y. and Noike, T. (1998). The Influence of pH and Ammonia
- 409 Concentration on the Methane Production in High-Solids Digestion Processes. WER,
- 410 **70**(5), 1075.
- 411 Lehne, G., Müller, A. and Schwedes, J. (2001). Mechanical Disintegration of Sewage
- 412 Sludge. Wat. Sci. Tech., 43(1), 19.

- 413 Li, Y.Y. and Noike T. (1992). Upgrading of Anaerobic Digestion of Waste Activated
- 414 Sludge by Thermal Pretreatment. *Wat. Sci. Tech.*, **26**(3-4), 857.
- 415 Liu, T. and Sung, S. (2001). Effect of Ammonia Inhibition in Thermophilic Anaerobic
- 416 Process. Proceedings of WEF/AWWA/CWEA Joint Residuals and Biosolids Management
- 417 *Conference*, San Diego, CA, February 21-24.
- 418 Kepp, U., Machenbach, I., Weisz, N. and Solheim, O.E. (2000). Enhanced Stabilization
- 419 of Sewage Sludge Through Thermal Hydrolysis: Three Years of Experience with Full-
- 420 scale Plant. Wat. Sci. Tech., 41(3), 213.
- 421 McCarty, P.L. and McKinney, R.E. (1961). Salt Toxicity in Anaerobic Digestion.
 422 J.W.P.C.F, 33, 399.
- 423 Richards, B.K., Cummings, R. J., Herndon, F.G. and Jewell, W.J. (1991). High-Solids
- 424 Anaerobic Fermentation of Sorghum and Cellulose. *Biomass and Bioenergy*, 1, 47.
- 425 Schafer, P.L., Farrell, J.B., Newman, G. and Vanderburgh, S. (2003). Advanced
- 426 Anaerobic Digestion Performance Comparisons. Proceedings of WEFTEC 03, Los
- 427 Angeles, CA, October 11-15.
- 428 Skiadas, I.V., Gavala, H.N., Lu, J and Ahring, B.K. (2005). Thermal pre-treatment of
- 429 primary and secondary sludge at 70 °°C prior to anaerobic digestion. Wat. Sci. Tech.,
- 430 52(1-2), 161.
- 431 Stephenson, R.J., Laliberte, S. and Elson, P. (2004). Use of a High-Pressure
- 432 Homogenizer to Pre-treat Municipal Biosolids: Introducing the $MicroSludge^{TM}$ Process.
- 433 Proceedings of *IWA 10th World Congress on Anaerobic Digestion*, Montreal, Canada,
- 434 August 29-September 2.





Figure 2. Total and Soluble COD Concentrations for CPAS and HCPAS

Figure 3. Total Volatile Solids Loading



Figure 4. Total Volatile Solids Destruction



Figure 5. Total and Soluble COD Removed



480 Table 1. Digester Feed Schedules, Total Solids and HRT

Digester	Feed Volume (L)	Periodicity	TS	HRT
_		(min)	(%)	(d)
Control MAD	0.63	60	2.5-4.5	15
THMAD	0.75	144	7-10	15
	0.94	144	7-10	12
Control TPAD Thermo Stage	0.95	60	3-4	5
Control TPAD Meso Stage	0.95	60	2-3	10
THTPAD Thermo Stage	0.95	60	7-10	5
_	1.18	60	6-8	4
	1.57	60	8-9	3
THTPAD Meso Stage	0.95	60	5.5-8.5	10
	1.18	60	4.5 -6	8
	1.57	60	6.5-7.5	6

482 Table 2. Operational Data During Steady State Operation of Pilot Digesters

Process	HRT	pН	Alkalinity	Total Volatile	Ammonia
	(Day)	(Unit)	(mg/L)	Acids (mg/L)	(mg/L)
Control MAD	15	7.2-7.6	3500-4500	200-500	1000-1500
THMAD	15	7.7-7.9	7200-9000	300-800	2200-2700
	12	7.3-7.7	5100-5600	400-900	2000-2200
Control TPAD	15	7.1-7.5	2900-3300	700-1400	1300-1500
Thermophilic					
Control TPAD	15	7.3-7.6	4500-4700	300-800	1500-2000
Mesophilic					
THTPAD	5	6.7-7.2	4400-5800	6700-8400	2000-2600
Thermophilic	4	6.9-7.3	4300-5400	4900-6400	1800-2200
	3	6.6-6.9	4200-5000	6700-8100	2300-2500
THTPAD	10	7.3-7.7	6000-7800	600-1200	1700-2300
Mesophilic	8	7.5-7.8	6200-7600	400-700	1700-2300
	6	7.3-7.7	6200-7300	400-1300	1700-2100

485 Table 3. Daily gas production per unit volume of digester capacity installed

Process	Gas Production
	m [°] CH ₄ /m [°] .d
Control MAD	0.8-1.2
THMAD	2-3
Control TPAD	1.1-1.7
THTPAD	2.8-4

Parameter	2005	THTPAD	Change
	Data	Trojections	(/0)
Dry Solids (Mg/d)	82	82	-
VS Destruction (%)	53	65	+23
Truck Loads	216	120	-44
Biogas (m^3/d)	44000	54000	+23
Energy in biogas (MWh)	270	350	+30

490 Table 4. SEP Average Daily Biosolids Inventory